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## MANAGEMENT OF SHARKS IN NEW SOUTH WALES WATERS

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### Research Article

## Aerial Survey as a Tool to Estimate Abundance and Describe Distribution of a Carcharhinid Species, the Lemon Shark, *Negaprion brevirostris*

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Aerial survey provides an important tool to assess the abundance of both terrestrial and marine vertebrates. To date, limited work has tested the effectiveness of this technique to estimate the abundance of smaller shark species. In Bimini, Bahamas, the lemon shark (*Negaprion brevirostris*) shows high site fidelity to a shallow sandy lagoon, providing an ideal test species to determine the effectiveness of localised aerial survey techniques for a Carcharhinid species in shallow subtropical waters. Between September 2007 and September 2008, visual surveys were conducted from light aircraft following defined transects ranging in length between 8.8 and 4.4 km. Count results were corrected for "availability", "perception", and "survey intensity" to provide unbiased abundance estimates. The abundance of lemon sharks was greatest in the central area of the lagoon during high tide, with a change in abundance distribution to the east and western regions of the lagoon with low tide. Mean abundance of sharks was estimated at 49 ( $\pm$ 8.6) individuals, and monthly abundance was significantly positively correlated with mean water temperature. The successful implementation of the aerial survey technique highlighted the potential of further employment for shark abundance assessments in shallow coastal marine environments.

#### 1. Introduction

Aerial survey has been used as a tool to assess species abundance for both terrestrial and marine vertebrates, often where the remoteness or vastness of the survey area and the potentially low abundance of the study species render other techniques uneconomical [1–6]. In the marine environment, aerial survey has typically focused on air breathing marine mammals and reptiles [2, 7–13] because these taxa are regularly visible at the surface. Through aerial surveys, it has been possible to quantify the abundance of marine creatures in remote locations. A few examples of these are dense concentrations of narwhal (*Monodon monoceros*) in the offshore pack ice of Baffin Bay, West Greenland [14] and the seasonal distribution of crabeater seals (*Lobodon carcinophagus*) in the pack ice of Antarctica [15]. In addition, aerial surveys have revealed unique insights into marine creatures, for example, the specific birthing location of an endangered western North Atlantic right whale (*Eubalaena glacialis*) [16] and mass aggregations of whale sharks (*Rhincodon typus*), numbering up to 420 individuals, previously not witnessed [17].

For sharks, aerial survey has been largely limited to the large filter feeding species, whale [18–20] and basking sharks (*Cetorhinus maximus*) [21, 22], as both these species spend long periods of time feeding near the surface, and their large size make them highly visible. In contrast to the large filter feeding species, most coastal shark species spend relatively little time at the surface, instead remaining near the seabed [23] or undertaking variable vertical diving profiles [24, 25]. These behaviours make coastal species generally unsuitable for aerial survey; hence, to date limited aerial survey of coastal sharks has been undertaken. Initially, incidental accounts of sharks observed during aerial surveys designed to survey turtles and marine mammals were published for seasonal occurrence of hammerheads (*Sphyrna* spp.) in Cape Canaveral, Florida [26], and the distribution of various shark species in the northeast United States [27]. For surveys intended to focus on coastal sharks, Gruber et al. [28] conducted experimental surveys of the Bimini lagoon lemon sharks (*Negaprion brevirostris*) from an ultralight aircraft; then Reyier et al. [29] opportunistically surveyed juvenile lemon sharks aggregating along the Cape Canaveral shoreline, Florida. Most recently aerial surveys were used to identify a possible inshore white shark (*Carcharodon carcharias*) nursery in Algoa Bay, South Africa [30].

The lemon shark (Negaprion brevirostris; Poey 1868) is classified as a large coastal shark species [31]. Lemon sharks are one of the larger Carcharhinid species, reaching a maximum length of ~260 cm Pre-Caudal Length (PCL), with both females and males reaching sexual maturity at around 12 years of age and ~166 cm PCL [32, 33]. Three distinct ontogenetic life stages are recognised for lemon sharks; nursery-bound juvenile (~45-80 cm PCL), subadult (~81-165 cm PCL), and adult/mature (~166+ cm PCL) [23, 34, 35]. They most commonly inhabit shallow subtropical waters around coral reefs, mangroves, seagrass beds, enclosed bays, sounds and river mouths [35]. Their diet consists mainly of bony fish and crustaceans [36-41]. Lemon sharks are viviparous, with a fecundity of 4-17 individuals [42]. Females display a biennial breeding cycle [42] and a high level of philopatry [43]; therefore the environmental health and state of specific nursery sites are important for breeding ecology. Lemon sharks are commercially targeted [31, 44], and their conservation status is currently listed as "Near Threatened" by the International Union for Conservation of Nature (IUCN) Red List, largely due to habitat degradation of nursery areas [45].

With increasing concern over the global status of shark populations [46] coupled with the inherent difficulties of studying large mobile marine predators, new techniques, which provide reliable species abundance estimates, are required. Aerial survey of coastal shark species in shallow subtropical waters, where water clarity and visibility are generally high, may provide such a tool. Through the use of small, low flying aircraft and visual survey, the aims of the study were to (1) estimate subadult lemon shark abundance in Bimini lagoon, (2) delineate the spatial distribution of subadult lemon shark abundance in the lagoon and variation relative to tidal state, (3) determine seasonal variations in subadult lemon shark abundance relative to water temperature, and (4) assess the effectiveness of aerial surveys for providing coastal shark abundance estimates.

#### 2. Materials and Methods

2.1. Study Site. The Bimini islands, Bahamas  $(25^{\circ}44' \text{ N}, 79^{\circ}16' \text{ W})$ , are located on the western edge of the Great Bahama Bank, adjacent to the deep waters of the Gulf Stream (Figure 1). The two main islands, North and South Bimini,

lie on either side of a shallow sandy lagoon measuring approximately  $25 \text{ km}^2$  and  $\sim 1 \text{ m}$  average depth. Lemon sharks have known nursery habitat in the Bimini islands, with the home range of individuals increasing with increasing size of animal. For the first three years of life, juveniles inhabit specific near shore mangrove primary nurseries, increasing their home range exponentially with each year of growth [37, 38, 47, 48]. At around four years of age they expand their movements and use the central lagoon for the duration of their subadult life stage [23, 28, 49]. The Bimini lagoon is a shallow (~1m), clear water environment, with a mostly light coloured sandy bottom (Figure 1), providing an ideal aerial viewing situation for detecting the darker lemon sharks swimming near or resting on the bottom (see [18]). Preliminary aerial observations at Bimini demonstrated the feasibility of this technique to provide realistic estimates of lemon shark abundance if conducted in a systematic way [28]. Due to the light coloured substratum and known subadult lemon shark site fidelity, the central lagoon was selected as the study site to be covered by the aerial survey.

2.2. Aerial Survey. The aerial survey design adopted stratified block sampling protocol following standard procedures [1, 3, 18, 50, 51]. The study area was divided into blocks using ArcGIS 10 (Figure 2), by dividing the central lagoon into near equal-sized sections with marker buoys, and then the rest of the survey area was divided into similar-sized blocks using existing landmarks. During each aerial survey, defined as one sampling event, a total of 10 transects were flown providing two passes for each block, one pass on the western edge when travelling north and one pass on the eastern edge when travelling south (Figure 2). Due to transects being positioned on the survey block edges, some of the survey tracked the lagoon shoreline. This would have potentially positively influenced abundance estimates if the subadult lemon sharks displayed a level of shoreline attachment/site fidelity. However, extensive acoustic tracking studies revealed no such patterns in their distribution, showing wide lemon shark distribution across the entire lagoon area [23, 52]. Therefore, no such positive influence was expected, and it was concluded that survey block edge transects would provide an unbiased sample of the survey region. The survey area covered during every sampling event, within each defined block, was then calculated from each strip transect (see below). Total survey time for each sampling event ranged from 27 to 32 minutes (mean  $\pm$  SE) 30.62  $\pm$  0.5. Aerial surveys were only conducted in wind conditions below Beaufort scale 3 (12–19 kmh<sup>-1</sup>): when cloud cover was <30% and between 10:00 and 14:00, as close to 12:00 as possible and when the sun is at its highest point to minimise bias associated with these environmental conditions [3, 9, 18, 50, 53].

The aerial survey team consisted of four members: a pilot, two observers, and a survey monitor. Both observers were located on the starboard side of the aircraft adopting the dual observer technique [3]. The first observer was located in the front right seat facing right and forward ahead of the wing and the second observer in the rear of the aircraft facing right and backward behind the wing, eliminating the potential of



FIGURE 1: Aerial Photograph of the Bimini Islands with North Bimini above and South Bimini below encasing the study site, the shallow, sandy central lagoon.



FIGURE 2: Aerial survey transect flight path depicted by the line. Arrows indicate the direction of the flight path. A–S represent stratified block sampling blocks.

visual obstruction in low winged aircraft. The two observers identified shark species, undertook counts for each block and recorded data on predefined survey sheets independently with no in-flight communication. The same two observers were used for all surveys to standardise potential observer bias. Both observers were trained and experienced in shark identification. Following the completion of a sampling event, observer survey data were examined, and the maximum count of sharks for each survey block, that is, the largest number of sharks observed by both or either of observers per block, was recorded. The survey monitor, located behind the pilot, recorded the GPS tracks of each transect to ensure accurate repeatability. Additionally, time of takeoff 3

and landing, start and end time of each transect, altitude, cloud cover, wind speed, and wind direction were recorded.

Throughout the study period, four different aircrafts were used to undertake aerial surveys (Cessna 172, a Beechcraft 35 Bonanza, a Piper Pa-28 Archer, and a Piper PA-31-350 Navajo Chieftain). Each aircraft was similar in design, and therefore all surveys were conducted with team members located in the same positions and under the same flight conditions. For all sampling events (n = 8), aircrafts were flown at an altitude of 100 m and a groundspeed of 185 kmh<sup>-1</sup>. We assumed that there was no movement of individual sharks between blocks during the duration of each aerial survey. This was based on the published average lemon shark swimming speed of 0.57 ms<sup>-1</sup> [54]. The abundance estimate (N) was calculated for each block using Rowat et al. [18] equation as follows

$$N = \frac{(C \times \text{ACF} \times \text{PCF})}{\text{SI}},$$
 (1)

where *C* is the lemon shark count, ACF is the availability correction factor, PCF is the perception correction factor, and SI is the survey intensity (see definitions below). Finally, the sum of all block abundance estimates provided an abundance estimate for the total survey area.

2.3. Availability Correction Factor (ACF). The availability correction factor (ACF), also known as availability bias, accounts for potential sharks that were within a survey block but were not visible and therefore not counted. For animals in the marine environment, visual counts of individuals are normally restricted by water turbidity, water/animal depth, and diving behaviour [1–3, 18, 19]. In Bimini, the maximum water depth within the survey area was <4 m; thus water depth and diving behaviour were not considered to be variables affecting lemon shark availability for visual detection. Surveys were also restricted to calm conditions (wind < 12–19 kmh<sup>-1</sup>) with little to no turbidity; therefore turbidity effects could be discounted. Thus, in this study, all sharks residing within the surveyed area during all surveys would be available to be counted.

2.4. Perception Correction Factor (PCF). The perception correction factor (PCF), also known as perception bias, accounts for the number of individual sharks not counted by the observers due to factors such as glare, fatigue, or inattention [2, 3]. For aerial survey, the two greatest potential sources for perception error are aircraft altitude and observer error. For a given aircraft altitude, it may not be possible to see an individual of a given size, with increasing altitude decreasing the ability of observers to count small individuals. Marsh and Sinclair [3] reported that a ratio of "altitude: minimum animal size" of 274:1 enabled reliable detections. In the present study an aircraft altitude of 100 m was maintained, and the minimum animal target size was ~1 m total length, providing a ratio of 100:1, well within the limits of reliable detection. Smaller juvenile lemon sharks, <1 m total length, would have been difficult to distinguish and tend to avoid the exposed lagoon areas, segregating due to predation threat from larger conspecifics [52]. Therefore, identified individuals in the survey area were assumed to be subadults. PCF correction for altitude was therefore not required. Observers may simply miss or fail to identify individual animals within a survey area for a number of reasons that can vary between observers [3]. For this reason, two independent observers were used for each survey, and we adopted Marsh and Sinclair's [3] dual observer bias equation to correct for perception bias

$$PCF = \frac{\left(S_m + b\right)\left(S_r + b\right)}{b\left(S_m + S_r + b\right)},$$
(2)

where " $S_m$ " is the number of sharks seen by observer 1 only, " $S_r$ " is the number of sharks seen by observer 2 only, and "b" is the number of sharks seen by both observers.

*2.5. Survey Intensity (SI).* Survey intensity (SI) is the proportion of the total survey area sampled by the aerial survey and was calculated as follows [18]

$$SI = \frac{\text{area of coverage}}{\text{area of block}}.$$
 (3)

The SI value accounts for sharks that are potentially present in the area of the block not covered by the visual swath of the flight path. SI values were calculated for each block using ArcGIS 10. First, individual polygons were created for each of the aerial survey blocks to calculate the total area of each block. The effective visual transect width for each block during an aerial survey was established for the standard survey altitude and speed (100 m and 185 kmh<sup>-1</sup>, resp.) using submerged markers ( $100 \times 40$  cm). The submerged markers were made from plywood and positioned in the lagoon at 25 m distance intervals from a known location on the flight path. The number of markers visible was counted on a presurvey test flight to verify the visual transect width. As the test flight and all subsequent survey flights were conducted within the predefined survey conditions and due to the uniform characteristics of the survey area, it was assumed that visual transect width would remain relatively constant between surveys. ArcGIS 10 was used to calculate the survey area, visually covered, by "buffering" (ArcGIS tool) the flight path with the calculated visual transect width. Area of coverage was then obtained by the "union" (ArcGIS tool) of the block polygons, with the visual transect width buffer. By this union process, each block polygon was divided into three subsequent polygons; two of which represented the visual area of coverage for the north and south line transects. The combined areas of the two isolated visual coverage polygons provided the area of coverage.

2.6. Lemon Shark Abundance. For each survey, lemon shark abundance for each block was calculated and then summed to give a total abundance estimate for the entire survey area. The calculated abundance for each block was then divided by the area of that block to give an abundance value expressed as *NB*/km<sup>2</sup>. Calculated *NB*/km<sup>2</sup> values for all survey events were combined for each block and then divided by the number of surveys conducted to give mean unit effort (UE) for lemon shark abundance per block. This was then mapped in ArcGIS

10 to provide a spatial depiction of the results. Aerial survey events were then divided into low- and high-tide counts, and the above calculations were repeated. Percent change of the raw count data from high to low tide was calculated for each survey block and mapped in ArcGIS 10. The Chi-squared ( $\chi^2$ ) analysis was conducted on the raw count data to test if there was significant variation (P < 0.05) for survey block abundance would be equal.

2.7. Water Temperature Monitoring. Mean monthly water temperatures (°C) were obtained from 15 thermochron (iButton) temperature loggers deployed across study area. Spearman's rank correlation was used to test for a significant relationship (P < 0.05) between total monthly abundance estimates and mean monthly water temperatures, for all surveys, and then repeated for low tide surveys only to eliminate any potential influence of tidal state over lemon shark abundance.

#### 3. Results

*3.1. Survey Summary.* A total of eight aerial surveys were conducted between September 2007 and September 2008, two at high tide and six at low tide. In all surveys, a total of 212 sharks of three species, lemon (145), nurse (*Ginglymostoma cirratum*; 42), and blacktip (*Carcharhinus limbatus*; 25) sharks, were recorded in the study area. Two additional species, tiger (*Galeocerdo cuvier*) and bull (*Carcharhinus leucas*) sharks, were identified from the air, but they were only present outside of the survey area and therefore not counted.

3.2. Correction Factors and Survey Intensity. For the availability correction factor (ACF), no factors were limiting the availability of sharks present in the surveyed area; therefore an ACF of 1 was applied. Perception correction factor (PCF) values ranged between surveys from 1.13 to 1.4 (mean =  $1.28 \pm 0.03$ ; Table 1). The visual transect width, calculated from the submerged markers, was found to be 350 m for both observers. This distance may appear to be large for an observational altitude of 100 m, but it was facilitated by the shallow waters, uniform light sandy substrate, and relative high contrast of the target species. Block survey intensity (SI) ranged from 0.332 to 0.885 (mean =  $0.521 \pm 0.029$ ; Table 2).

3.3. Lemon Shark Abundance Estimates. Mean lemon shark abundance for the entire survey period was 49 (±8.6) individuals, ranging from a minimum 16 sharks in March 2008 to a maximum of 80 sharks in September 2008. Lemon shark abundance was generally higher in the summer months relative to the winter, a trend present when considering all surveys, high and low tide, and low tide surveys only (Figure 3). Mean monthly lemon shark abundance estimates were significantly correlated with mean monthly water temperature both for all surveys (r = 0.78, n = 8, P < 0.05) and surveys conducted only at low tide (r = 0.88, n = 6, P < 0.05). For all eight aerial survey events, the highest abundance of lemon sharks was recorded in the most central blocks E–J

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Survey date	Total lemon sharks observed	b	S <sub>m</sub>	S <sub>r</sub>	PCF
Sep-07	24	9	7	8	1.26
Nov-07	18	6	4	8	1.30
Dec-07	10	3	4	3	1.40
Jan-08	16	6	5	5	1.26
Mar-08	6	2	2	2	1.33
May-08	12	6	3	3	1.13
Jul-08	26	9	8	9	1.31
Sep-08	33	13	10	10	1.23

TABLE 1: Perception correction factor (PCF) values for each aerial survey, where " $S_m$ " is the number of sharks seen by observer 1 only, " $S_r$ " is the number of sharks seen by observer 2 only, and "b" is the number of sharks seen by both observers.

TABLE 2: Area of block  $(m^2)$ , area of coverage  $(m^2)$  based on 350 m visual swath width, and survey intensity (SI) of each survey block A–S.

Survey block	Area of block (m <sup>2</sup> )	Area of coverage (m <sup>2</sup> )	SI
А	1,372,459	883,975	0.644
В	1,629,582	803,130	0.493
С	3,034,218	1,259,041	0.415
D	2,550,267	1,385,899	0.543
E	1,253,667	842,485	0.672
F	1,626,326	674,328	0.415
G	1,720,557	571,343	0.332
Н	1,471,679	681,739	0.463
Ι	1,427,818	658,329	0.461
J	1,657,296	897,071	0.541
Κ	745,701	660,251	0.885
L	2,864,207	1,910,109	0.667
М	3,225,809	1,871,598	0.580
Ν	2,019,770	955,687	0.473
0	1,407,947	667,035	0.474
Р	1,811,735	850,333	0.469
Q	2,253,654	1,009,107	0.448
R	1,450,254	663,801	0.458
S	2,038,271	942,918	0.463

and N–P (Figure 4(a)). The overall highest abundance was recorded in block F (6.17  $NB/km^2$ ) and the lowest in block Q (0.16  $NB/km^2$ ). The mean lemon shark abundance of all blocks was 1.15 ± 0.33  $NB/km^2$ .

The abundance distribution of lemon sharks varied between high and low tides, although a disproportionate number of low tide surveys relative to high-tide surveys were flown. From high to low tide, there was a change in abundance distribution from the centre of the lagoon with blocks A-B and G–K showing a decrease in abundance, while all other blocks showed an increase in abundance (Figures 4(b) and 5). From high to low tide surveys, mean lemon shark abundance (*NB*/km<sup>2</sup>) in block G and block I showed a highly significant decrease ( $\chi^2 = 30.1$ , d.f. = 1, P < 0.001), and block I showed a significant decrease ( $\chi^2 = 5.0$ , d.f. = 1, P < 0.05). Block F showed a highly significant increase ( $\chi^2 = 26.9$ , d.f. = 1, P < 0.001), and block O showed a significant increase  $(\chi^2 = 6.7, \text{ d.f.} = 1, P < 0.01)$ . Overall, two surveys were conducted on neap tide, five on half tides, and one on a spring tide. Mean tidal variation for all surveys was 0.63 m (range = 0.04–0.3 m). For surveys conducted at high tide, one was conducted on a neap tide and one on a half tide (mean variation = 0.54 ± SE 0.04). For surveys conducted at low tide, one was conducted on neap tide, four on half tide, and one on spring tide (mean variation = 0.65 ± SE 0.04). Given the low tidal variability between surveys, we assume that tidal type did not have a strong influence over abundance distribution.

#### 4. Discussion

This study represents the first successful focused employment of the aerial survey technique on a Carcharhinid species. This is useful not only for providing abundance estimates in a much shorter temporal scale than other methods, such as tag recapture estimates, but also for defining distribution without the potential biases. Sample biases can result from the use of longlines, where sharks can be attracted to the given areas by the bait [55], or active tracking, where the behaviour of the sharks can be altered by the presence of the boat [56]. In Bimini, aerial survey has shown the mean abundance of subadult lemons sharks to be 49 individuals. with the highest distribution of abundance in the centre of the lagoon and significantly influenced by tidal state. Observed abundance was also significantly positively correlated with water temperature, resulting in abundance variation across the course of the study. The aerial survey technique allowed for comprehensive abundance estimates, with equal sampling effort over a large area of secondary nursery, including otherwise hard to access areas due to very shallow water depths.

The block abundance of lemon sharks is highest in the central region of the lagoon (Figure 4(a)), which is the shallowest section of the lagoon and can even be exposed at spring low tide (personal observations 2002–2009; [28]). This is consistent with past life history data describing an ontogenetic shift from the near-shore mangrove fringed nurseries to the more open but shallow central lagoon area at the subadult life stage [23, 34, 48]. Despite disproportion in the surveying of the two tidal states (high n = 2 and low n = 6) the change in abundance distribution between tidal states is consistent with results of several acoustic tracking studies



FIGURE 3: Lemon shark total abundance (number of lemon sharks) estimates by date, derived from aerial survey counts. Line graph represents mean monthly water temperature ( $^{\circ}$ C); grey bars represent surveys conducted at low tide, and black bars represent surveys conducted at high tide.



FIGURE 4: (a) Mean lemon shark block abundance for all surveys conducted; (b) change in block lemon shark count per unit effort (UE) from high-tide to low-tide surveys. Dark grey blocks represent an increase in count per UE; light grey block represents a decrease in count per UE; patterned block represents no change in count per UE,  $\chi^2$  significant \*P < 0.05, \*\*P < 0.01, \*\*\*P < 0.001.



FIGURE 5: Percent change in mean lemon shark block counts (unit effort; UE) from high tide to low (positive) and low to high (negative) tide surveys;  $\chi^2$  significant *P* values denoted by  ${}^*P < 0.05$ ,  ${}^{**}P < 0.01$ ,  ${}^{***}P < 0.001$ .

[23, 28, 52, 55]. The spatial change in abundance distribution, with the central areas of the lagoon showing high abundance at high tide and low abundance at low tide (Figures 4(b) and 5), was probably the result of decreased water depth making the central area physically uninhabitable at the low tidal state. The significant tidal changes in abundance for block I and block O and highly significant changes for block F and block G are consistent with areas of the lagoon where habitability is greatly affected by tidal change. Block G and block I are particularly shallow areas that become almost dry during extreme low tides. Thus, sharks would not have been able to physically inhabit these areas at low tide. Block O is adjacent to block I, but it still retains a sufficient water depth at low tide to accommodate lemon sharks, with water depth progressively increasing to the east. Block F contains a large deep section ( $\sim$ 1.5 m) that at low tide is cut off from large predator risk by surrounding shallow water, but it remains deep enough for lemon sharks to inhabit. This potential lowtide refuge function, plus its adjacency to block G showing high abundance at high tide, is a probable driver for the highly significant increase in abundance between high and low tide.

Water depth appears to be defining the lemon shark's distribution, with the shallowest inhabitable areas being favoured. The Bimini lagoon area has been previously described as a secondary nursery for lemon sharks, indicating that it offers the inhabiting individuals some form of protection from potential predators [23, 28]. As the Bimini lagoon is generally homogenous with little structure present to offer physical protection, the shallow nature of the lagoon may offer protection through limited water depth. The central area is generally the shallowest; therefore, it is generally inaccessible to larger sharks that might pose a predation risk to smaller lemon sharks. As the tide falls, the central area becomes uninhabitable, forcing the lemon sharks into other areas that retain sufficient water depth. These other areas, towards the edges of the lagoon, then offer a shallow refuge as they have reduced in water depth with the falling tide. With a distribution influenced by predation risk, tidal

driven abundance distribution changes are expected, and they have been commonly documented in other primary and secondary nurseries for several shark species [49, 52, 57–61]. The results of this study support the description of the central lagoon as a secondary nursery for lemon sharks.

Monthly variations in abundance, independent of tidal variations, were recorded and would be logical for a life stage or species that is known to display large scale seasonal movements [18]. Chapman et al. [49] found that 51.3% of subadult lemon sharks genetically sampled in the Bimini lagoon were born into Bimini nurseries. Thus, 48.7% were born elsewhere, and due to the relative isolation of the Bimini islands from any other suitable lemon shark nursery sites (~115 km to the nearest mangrove fringed island), the subadult lemon sharks are able to undertake large movements at this life stage. The significant relationship found between abundance and mean monthly water temperature indicates that, as documented for other Carcharhinid species [62], water temperature could have been driving seasonal abundance variation in the Bimini lagoon. Lemon shark temperature preferences may have resulted in seasonal latitudinal migrations. It is possible that individuals could travel to the more southern Bahamian islands on the great Bahama Bank, such as Andros, where the water temperatures may be more favourable in the winter months. Around the Bimini islands themselves, the lemon sharks may have moved to the deeper and warmer waters on the edge of the adjacent Gulf Stream, a behaviour exhibited by other ectothermic shark species in response to temperature change [60, 63]; however this may result in a considerable increase in predation risk.

It is possible that decreased lemon shark metabolism in response to decreased water temperatures resulted in decreased activity [64–70]. Lemon sharks were more visible when in motion relative to the surrounding habitat (personal observation, 2007-2008); therefore, active individuals may have been more likely to be recorded than resting individuals. Thus, the recorded reduction in abundance correlated to water temperature reduction may have been partly exaggerated to due to the increased perception error in the winter months.

The aerial survey technique proved effective for establishing lemon shark abundance estimates for the Bimini lagoon. Sharks were easily visible from the aircraft, and the lemon shark was easily distinguished from the other species present, due to their signature fin assemblage, colour, and swimming motion. Perception correction factor (PCF) values were reasonable relative to past studies focused on other marine vertebrates [71]. Abundance distribution was easily established, and it was consistent with past telemetry studies both in general distribution and tidal driven distribution variation [23, 28, 37, 52]. The consistency of distribution results, coupled with the results of the extensive number of lemon shark acoustic telemetry studies conducted in the survey area, indicates that the stratified block sampling design provided a representative subsample of the total lemon shark abundance. Subadult lemon sharks have been shown to use the full extent of the survey area, with no strong site fidelity to areas within the survey blocks not covered by the survey transects [23, 52]. Thus, the application of the appropriate correction factors presented in this study would have been successful in extrapolating representative abundance estimates across the entire survey area.

This represented the first employment of the aerial survey technique for the abundance assessment of a Carcharhinid species. With global concern over declines in shark abundance resulting from target fisheries and bycatch [46, 72-74], it is imperative that new methods for assessing regional shark abundance have to be identified. The success of the aerial survey approach in Bimini highlighted the potential for further employment of this research technique for shark population assessments in similar habitats. It should be noted, however, that the survey design employed in this study was relatively basic as it was facilitated by the highly favourable conditions of this study site. For more challenging environments, a more complex survey design may be required to achieve reliable and representative abundance estimates (see Buckland [75]). The aerial survey technique would be proved effective for any area where sharks have a high level of site fidelity to clear shallow waters. For example, this technique would be particularly useful for shallow coastal beach areas that are commonly the boundary of human and shark interactions that result in both social and conservation issues.

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